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2 Prospects and limitations of prescribed burning as a management tool
3 in European grasslands

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21

21 **Abstract**

22

23 Grassland managers and scientists are increasingly interested in cost-effective alternative
24 ways of grassland biodiversity conservation. Prescribed burning is a promising management
25 tool which should be integrated in the planning of management efforts. In addition, small-
26 scale prescribed burning is an effective fire suppression strategy to decrease the serious
27 negative impacts of uncontrolled burnings on ecosystems and human life. Prescribed burning
28 forms an integral part of the North-American grassland management practice, while in
29 Europe it is rarely applied, despite the fact that uncontrolled burning occurs frequently in
30 some regions. Our goal was to evaluate the use of prescribed burning as a promising but
31 neglected management tool in European grasslands. We found that European studies on
32 prescribed burning of grasslands are scarce and we conclude that annual burning is usually
33 not an appropriate option for the conservation of species-rich grasslands. We reviewed
34 burning studies from North-America to identify findings which might be adapted to the
35 European grassland conservation strategy. In North-America, contrary to Europe, the
36 application of burning is fine tuned in terms of frequency and timing, and usually combined
37 with other restoration measures (grazing or seed sowing). Thus, we conclude that with the
38 application of carefully designed prescribed burning, multiple conservation goals, e.g.
39 invasion control and increasing of landscape-level heterogeneity, can be linked with an
40 effective fire suppression strategy. We emphasize that for the application of prescribed
41 burning in Europe, the general findings of carefully designed case studies should be
42 combined with the practical knowledge of conservation managers concerning the local
43 application circumstances to reach specific management objectives.

44

45 **Keywords:** Biomass; Ecosystem services; Fire; Grazing; Mowing; Prairie.

46

47 **Introduction**

48

49 Grasslands are of crucial importance in maintaining landscape-level biodiversity and are vital
50 elements of the historical landscape of Europe. The area and species richness of grasslands
51 have been in constant decline in many parts of Europe. Still existing grasslands are threatened
52 by the cessation of traditional management (Kahmen, Poschlod & Schreiber 2002); which
53 can lead to (i) litter accumulation (Ryser, Langenauer & Gigon 1995); (ii) increased fuel
54 loads resulting in regular wildfires (Baeza, Luís, Raventós & Escarre 2002), (iii)
55 encroachment of herbaceous competitors (Kahmen et al. 2002; Köhler, Gigon, Edwards,
56 Krüsi, Langenauer et al. 2005) or (iv) invasion of woody species (Hansson & Fogelfors
57 2000), each resulting in the decline of target grassland species in the long term.

58

59 Traditional grazing and mowing are no longer sustainable in many regions because of the
60 significant decrease in livestock-numbers and a reduced demand for forage. On the other
61 hand, grazing and mowing can have relatively high costs in grasslands with difficult
62 accessibility and located far from settlements (Köhler et al. 2005). Thus, conservation
63 managers and scientists are seeking less costly and labour-intensive approaches which can
64 also maintain grassland species richness and eliminate the negative consequences of
65 abandonment (Köhler et al. 2005; Liira, Issak, Jögar, Mändoja & Zobel 2009). It seems
66 important to test whether prescribed burning is an appropriate substitution of grazing and
67 mowing in European grasslands based on carefully designed evidence-based case studies.

68

69 For developing improved grassland management strategies, the evaluation of fire effects on
70 grassland structure and species composition is crucial. Prescribed burning studies can

3

71 contribute to the understanding of the ecological impacts of uncontrolled wildfires and arson,
72 which are present in many regions of Europe. According to recent climate change scenarios,
73 climate will be warmer and drier, which will increase the probability of wildfires, especially
74 in the Mediterranean region (Pausas 1999). Due to warmer and drier climate and increased
75 fuel loads caused by abandonment, the probability of wildfires will increase even in those
76 countries which are scarcely affected by wildfires recently. Thus, fire suppression strategies
77 against uncontrolled wildfires will need to be developed in the future (Castellnou, Kraus &
78 Miralles 2010). Application of carefully designed, small-scale prescribed burnings can be an
79 effective fire suppression strategy to mitigate the serious negative impacts of wildfires and
80 uncontrolled burning (Baeza, Luís, Raventós & Escarre 2002). To meet long-term resource
81 management and conservation goals, the application of prescribed burning under specified
82 fuel and weather conditions is necessary (Castellnou et al. 2010).

83

84 Our goal was to evaluate the results of European attempts to use prescribed burning in
85 grassland management, and assess whether the targeted objectives were achieved. We
86 discussed burning studies from North-America as a reference system to identify which
87 elements of fire management can be adapted to the European grassland conservation strategy.

88

89

90 **Material and methods**

91

92 We obtained information from three levels: (i) a literature search of scientific electronic
93 databases, (ii) a search in professional networks and (iii) direct contact with conservation
94 experts. First, we collected papers by searching in the database ISI Web of Knowledge for the
95 period 1975-2012, using the keywords ‘prescribed fire’ OR ‘prescribed burn*’ AND

4

96 'grassland' which yielded 480 hits (last accessed 18/12/2012). The terms 'Europe' and
97 'North-America' were omitted from the search keywords as suggested by the systematic
98 review protocol of Pullin *et al.* (2006), because relevant studies that do not mention these
99 terms may have been missed. We restricted the results to (i) European countries yielding 26
100 hits and (ii) North-American countries yielding 397 hits. The significant bias between
101 European and North-American studies on the topic did not allow us to execute a meta-
102 analysis.

103

104 The study inclusion criteria were the following:

- 105 • Relevant subjects: all types of grasslands in Europe or North-America; shrublands,
106 marshlands and heathlands were not considered
- 107 • Types of treatment: prescribed burning; excluding wildfires and arson
- 108 • Types of comparator: comparison with similar grassland plots which were not burned;
109 at least unmanaged (control), but if other management types were studied, we
110 considered them as well (e.g. mowing, mulching or grazing)
- 111 • Types of variables: species richness or abundance of any taxa or functional groups
112 (e.g. woody species, tall herbs) and ecosystem properties (e.g. amount of litter)

113

114 Out of the 26 results found for European countries, none matched these criteria. Thus, we
115 started an additional search using the keywords 'fire' OR 'burn*' AND 'grassland' (resulting
116 in 3,833 studies in total) focusing on European countries, which yielded 595 results. All the
117 595 studies were scanned at title, abstract and full-text level and finally 8 studies matched the
118 selection criteria.

119

120 Second, we searched altogether 18 volumes of *International Forest Fire News*, which is not
121 indexed in scientific electronic databases ([http://www.fire.uni-](http://www.fire.uni-freiburg.de/iffn/iffn_online.htm)
122 [freiburg.de/iffn/iffn_online.htm](http://www.fire.uni-freiburg.de/iffn/iffn_online.htm)) and also the website of the Eurasian Fire in Nature
123 Conservation Network (<http://www.fire.uni-freiburg.de/programmes/natcon/natcon.htm>).
124 This search resulted in further three papers matching the study inclusion criteria. Finally, to
125 have a clearer view of the current European situation, we contacted several grassland
126 specialists across Europe to gain information concerning: (i) the regulation of burning by law,
127 (ii) the occurrence and frequency of wildfires and arsons in grasslands and (iii) the
128 possibilities and limitations of the use of prescribed burning in European countries. We
129 distributed questionnaires via e-mailing and through the mailing lists of the European Dry
130 Grassland Group and European Vegetation Survey (1,600 people), and we gained
131 information from 49 colleagues from 19 countries.

132

133 **Published evidence on grassland conservation using prescribed burning in Europe**

134

135 Altogether we found eleven European studies meeting the selection criteria (Table 1). In most
136 of the studies dormant-season burning was applied on an annual basis with a valuable long-
137 term monitoring (up to 28 years, Wahlman & Milberg 2002). Generally no data about pre-
138 burn species composition was given, only a brief description. Only a few studies evaluated
139 effects of burning on animals. Most studies were comparative experiments of potential
140 alternatives (e.g. burning or mulching) for traditional grazing or mowing, thus they did not
141 focus on the application of burning. Burning was chosen as a labour- and cost-effective
142 method compared to other management measures. In these studies burning was not combined
143 with any other management or post-fire rehabilitation.

144

145 The European studies concluded that annual burning alone is not appropriate to maintain the
146 desirable structure and species richness of the studied grasslands. In the long term, species
147 richness usually decreased in the burning treatment compared to grazing or mowing
148 treatments. Burning led to the increased dominance of competitor species like *Brachypodium*
149 *pinnatum* (Kahmen et al. 2002; Köhler et al. 2005), and resulted in an untargeted species
150 composition, similar to that of abandoned plots. The reason why burning proved
151 inappropriate in these studies might be because annual burning was applied for many years,
152 and the vegetation did not have enough time to regenerate between burns.

153

154 Only minor and not always significant advantages of burning were identified in the reviewed
155 papers. Although burning did not result in the targeted species composition, it promoted some
156 rare or endangered species of dry limestone grasslands like *Aster amellus*, *Gentianella ciliata*
157 or *Thesium bavarum* (Köhler et al. 2005). The elimination of litter layer (e.g. Liira et al.
158 2009; Ryser et al. 1995) and the delay of woody encroachment were also mentioned as
159 positive effects (Moog, Poschlod, Kahmen & Schreiber 2005; Page & Goldammer 2004).
160 Promising examples about the use of prescribed burning in the management of steppic
161 grasslands on viticulture terraces were published by Page & Goldammer (2004) and Rietze
162 (2009) (Table 1).

163

164 **The use of burning in European grasslands based on a questionnaire survey**

165

166 We received answers to our questionnaire from 49 grassland experts from Austria, Bulgaria,
167 Czech Republic, Estonia, France, Germany, Greece, Hungary, the Netherlands, Poland,
168 Portugal, Romania, Russia, Slovakia, Slovenia, Spain, Ukraine and the United Kingdom. In
169 the following, we refer to the results of the questionnaire survey by indicating country names.

7

170 Based on the questionnaire survey, burning was a traditional grassland management tool, to
171 improve forage quality, reduce woody encroachment and litter accumulation in many
172 countries (Austria, Czech Republic, Estonia, Greece, Hungary, Poland, Russia, Slovakia).
173 Recently, traditional fire use has disappeared from many countries because of intensification
174 of agriculture and socio-economic changes (Castellnou et al. 2010). Traditional burning
175 practices and the traditional ecological knowledge on grassland burning might hold great
176 potential for planning current grassland management. There is very little written information
177 on traditional burning practices, thus, further historical and ethnographic research is needed
178 to improve our knowledge on this topic (Castellnou et al. 2010).

179

180 Illegal, uncontrolled burning is practiced nowadays in extensive areas of Central-, Southern-
181 and Eastern-European countries, posing serious conservation and socio-economic problems
182 (Romania, Hungary, Bulgaria and Ukraine). There are several motives for setting fires
183 illegally, such as: (i) the improvement of pastures in mountain areas (Greece, France or
184 Romania); (ii) to gain Natura 2000 subsidies without labour-intensive management,
185 especially in lowland hay-meadows (Romania) or (iii) fires are set just for “fun” and
186 vandalism (Hungary, Romania and Ukraine). Given the unpredictable and often negative
187 impacts of uncontrolled fires, even prescribed burning is prohibited in most of the European
188 countries, to mitigate air pollution (Austria) and/or to protect human life and property
189 (Greece). There are some countries where prescribed burning is permitted with strict
190 regulations regarding the timing and extension of prescribed fires and the appropriate fuel
191 and weather conditions for burning (Germany, France, Spain, Portugal, the United Kingdom,
192 the Netherlands and Slovenia). There are detailed codes and training for professional teams
193 who apply prescribed burning mainly for heathland and shrubland management and fire
194 hazard reduction (Castellnou et al. 2010). In a few countries, prescribed burning is included

195 in the management of protected areas (e.g. in France or Portugal), but only a few studies are
196 available in English and their majority focuses on shrublands and heathlands.

197

198 **Key findings of North-American case studies**

199

200 Historically, fire had a higher impact shaping grasslands in North-America than in Europe.
201 As suggested by a global simulation model (Bond, Woodward & Midgley 2005), North-
202 American grasslands are more fire-prone than European ones. North-American grasslands are
203 mainly characterized by the more fire-adapted C4 grasses, while in Europe C4 grasslands are
204 not typical. Thus, fire was likely not a factor in the evolutionary history of many grassland
205 species from Europe. Another difference in fire regimes between the two continents is that in
206 North-America, fuel loads were more continuous than in Europe until recent times. In Europe
207 urbanization processes (creating fire breaks by linear infrastructures and settlements) started
208 much earlier than in North America, which decreased the extent and magnitude of wildfires.

209

210 In North-America prescribed burning is frequently used in grassland management programs,
211 and it is indicated by the large number of studies on this topic. In North-America, burning is
212 not only used as a substitutive tool for other management measures, but often combined with
213 other tools (grazing or seed sowing) and the overall aim of burning is often the reintroduction
214 of natural disturbance regimes (MacDougall & Turkington 2007). Prescribed burning is used
215 as a management tool in various North-American grassland types, mainly in tall-grass and
216 short-grass prairies and Mediterranean annual grasslands. In the following section, we
217 summarize the most important experiences of North-American burning practices which
218 could, at least partly, be adapted to European grasslands.

219

220 *Timing of burning.* In North-America both dormant- and growing-season burning are applied
221 to achieve management goals considering the phenology (e.g. germination, seed set and
222 dispersal) of target and unwanted species (Pyke, Brooks & D'Antonio 2010). Dormant-
223 season burning is most effective for the reduction of accumulated litter (Rowe 2010). Natural
224 fire regimes are best simulated by growing-season mid-July burns, at the peak of lightning-
225 season (Howe 1994). Most prescribed burning is applied in the spring in the USA, but
226 summer burning is also applied (Fuhlendorf, Engle, Kerby & Hamilton 2009). Summer fires
227 can be used (i) to suppress unwanted species in a phenological state most susceptible to fire;
228 or (ii) to give advantage to early-growing species which can regenerate after fire in autumn
229 (Howe 1994). Summer fires can cause serious damages in grassland species, as most plant
230 and animal species are active in this period (Fuhlendorf et al. 2009). Besides burning season,
231 fire effects also depend on fuel moisture and weather conditions (Twidwell, Fuhlendorf,
232 Engle & Taylor 2012).

233

234 *Frequency of burning.* To mimic natural disturbance regimes and maintain grassland
235 biodiversity, burning every 2-3 years is recommended in tallgrass prairies (Fuhlendorf et al.
236 2009). This interval resembles most the natural wildfire regimes required for the regeneration
237 of grasslands (Rowe 2010). To control invasive species, high-frequency burning in several
238 consecutive years is needed. Repeated burning may prevent the regeneration of the invasive
239 species from vegetative buds or seed bank, and burning should be repeated until the seed
240 bank of the invasive species is destroyed and there is a low risk of re-colonization (Alexander
241 & D'Antonio 2003, Pyke et al. 2010).

242

243 *Combination of grazing and burning – patch-burning.* Fire and grazing interact through
244 positive and negative feedbacks resulting in a shifting spatial and temporal mosaic (fire-

245 grazing model; Fuhlendorf & Engle 2001). The model is based on the principle that free-
246 ranging grazers preferentially select recently burned patches with high-quality forage for
247 grazing. Grazers rarely choose patches that have not been burnt for several years. This leads
248 to litter and biomass accumulation, increased fuel loads and a higher probability of wildfires
249 there. A conservation effort can be fulfilled by the application of patch-burning management
250 to mimic natural disturbance regimes and to improve spatio-temporal heterogeneity of
251 grasslands (Fuhlendorf et al. 2001). Within a large area, burning is applied in patches, each
252 patch is being burnt periodically, e.g. once three years to leave time for grassland
253 regeneration to the pre-fire state. Patch-burning management has several advantages
254 compared to homogenous burning: (i) The co-existence of various fire regimes can maximize
255 species richness (Parr & Andersen 2006). (ii) The increased landscape-scale heterogeneity
256 promotes the coexistence of species with different habitat requirements. (iii) Grazing animals
257 can freely select patches with the best forage quality. (iv) Patch-burning can help to suppress
258 large wildfires by creating heterogeneous fuel structure where low-fuel patches can act as fire
259 breaks (Hobbs 1996).

260

261 *The use of burning for invasion control.* Burning is a more natural measure for invasion
262 control than the application of herbicides, which can persist in the soil and can be detrimental
263 to grassland species (DiTomaso 2000). Burning can be used for invasion control in cases
264 when the phenology of invasive and target native species is different or they are differently
265 adapted to fire (MacDougall & Turkington 2007, Pyke et al. 2010). Timing of burning plays a
266 crucial role, as inappropriately timed burning can even facilitate invasion in arid and semiarid
267 ecosystems (Keeley 2006). Burning can increase the effectiveness of herbicides providing a
268 better contact between the herbicide and the plant by removing litter (DiTomaso 2000). There
269 are promising examples for the use of prescribed burning in the control of *Taeniatherum*

270 *caput-medusae* (Davies & Sheley 2011) or *Lespedeza cuneata* (Cummings et al. 2007).
271 Combination of burning and grazing can also be used to control invasive plants. After fire,
272 unpalatable invasive plants allocate most of their energy to regeneration and less energy to
273 defensive organs and secondary metabolites and therefore they can be more effectively
274 suppressed by grazing (for *Lespedeza cuneata*; Cummings, Fuhlendorf & Engle 2007).

275

276 *Post-fire rehabilitation techniques.* These can be used to improve grassland recovery and
277 mitigate unwanted effects of burning on grassland species. To prevent soil erosion of burned
278 sites seeding of sterile and non-persistent cereal grains (nurse crop) can be applied (Keeley
279 2006). A more effective way of post-fire rehabilitation is mulching or transfer of plant
280 material, which can reduce erosion, but at the same time, propagules of target species can be
281 introduced to the site (Kiehl, Kirmer, Donath, Rasran & Hölzel 2010).

282

283 **What can be learned from European and North-American studies?**

284

285 Besides the increasing interest for alternative grassland management measures, only a few
286 studies address the applicability of prescribed burning in European grasslands. An important
287 reason for the limited number of European studies is that due to legislative limits in most
288 countries, evaluation of prescribed burning experiments is difficult or even impossible.
289 European publications on prescribed burning of grasslands mainly used a simplistic approach,
290 i.e. yearly burning of the entire grassland site for many years. On the contrary, in North-
291 America, prescribed burning is frequently and successfully used in grassland management
292 programs, indicated also by the huge number of studies on the topic. There is a need for
293 focused case studies to test whether the well-developed North-American burning regimes can
294 be adapted to the European grassland conservation strategy.

295

296 Given the differences in history, climate and composition of grasslands in the two continents,
297 the elements of North-American burning practice can only partly be applied in Europe. A
298 major difference is that in North-America, more fire-prone C4 grasslands are typical, while
299 European grasslands are mainly characterised by C3 grasses. Thus, as a first step, North-
300 American burning regimes should be evaluated to determine in which European grasslands
301 prescribed burning can be a proper management option. European studies on prescribed
302 burning are available mostly from dry and mesophilous grasslands, where too frequent
303 (annual) burning proved to be an inappropriate method. Thus, we cannot draw general
304 conclusions for the proper management of these grassland types. Based on the identified
305 failures and successes of the reviewed studies, the most promising management objectives of
306 prescribed burning experiments could be the following.

307

308 *Reducing accumulated biomass.* Both European and North-American studies proved that
309 dormant-season burning can effectively remove accumulated biomass from abandoned
310 grasslands (Ryser et al. 1995; Rowe 2010). Based on these findings, prescribed burning
311 should be tested on sites, where management by grazing or mowing is not feasible, like in
312 limestone grasslands (Ryser et al. 1995). Besides effective biomass removal, burning in
313 abandoned grasslands can result in untargeted species composition if applied too frequently,
314 as it was found in most European studies. Thus, proper fire return periods should be tested in
315 various grassland types and also fine-tuned to site characteristics (e.g. the rate of litter
316 accumulation or the presence of noxious competitor species in the vegetation). Fire return
317 periods applied in the more fire-prone tall-grass prairies (2-3 years, Fuhlendorf et al. 2009)

13

318 suggest that at least three years may be appropriate in European grasslands because they are
319 evolutionary less adapted to fire than North-American ones.

320

321 *Supporting target species by burning.* Some European studies mentioned positive effects of
322 burning on several rare or protected species. Fire promoted some limestone grassland species
323 probably by creating suitable germination microsites (Köhler et al. 2005). Prescribed burning
324 can also favour xerophilous target species by providing warmer and drier microclimate in
325 steppic grasslands on abandoned vineyards (Page & Goldammer 2004). Focused case studies
326 on certain target species could be integrated in future conservation actions. However, based
327 on North-American experiences, burning is not recommended at sites, where remnant
328 populations of endangered species are present (MacDougall & Turkington 2007).

329

330 *Management of open landscapes.* Several European studies found that prescribed burning can
331 help in the maintenance of open landscapes by the prevention of woody encroachment (Page
332 & Goldammer 2004). In extended open landscapes, like Central- and Eastern European
333 steppes, the introduction of patch-burning management can increase landscape-level
334 heterogeneity. Based on North-American experiences, combination of fire and grazing can
335 provide patches characterised by different amounts of green biomass and litter (Fuhlendorf &
336 Engle 2001). The increased structural and functional diversity can promote the coexistence of
337 species with different habitat requirements. In extent grassland areas, prescribed burning can
338 also be a proper tool for preventing extent and uncontrolled wildfires and accordingly it can
339 contribute to the protection of personal safety and private property (Baeza et al. 2002).

340

341 *Invasion control.* Beside of the serious conservation problems posed by invasive species, in
342 Europe the application of fire against invasives has not been studied yet. In North-America,
343 carefully designed prescribed burning is effectively used against several invasive species. For
344 the application of prescribed burning in invasion control, the followings should be
345 considered: (i) Based on North-American studies, growing-season fires can be the most
346 effective in the suppression of invasive species. For appropriate timing, the most susceptible
347 period of the given invasive species should be identified. (ii) Since growing-season fire can
348 have detrimental effects on most grassland species, invasion control by prescribed burning
349 should be first tested in degraded grasslands to avoid damaging populations of rare species.
350 (iii) To achieve long-term results, burning should be repeated until the invasive species
351 disappears both from the aboveground vegetation and the seed bank. (iv) For the recovery of
352 natural grassland vegetation, post-fire rehabilitation by sowing seeds of native grasses is
353 necessary. (v) Prescribed burning could also increase the effectiveness of other invasion
354 control methods, like grazing or herbicide application, thus, complex methods should also be
355 tested.

356 We pointed out that prescribed burning of grasslands should be integrated in the European
357 nature conservation practice. However, given the limited number of case studies in Europe;
358 further habitat-specific experiments are needed to find specific management objectives and
359 application circumstances.

360

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376

377 **References**

378

379 Alexander, J.M., & D'Antonio, C.M. (2003). Seed bank dynamics of French Broom in coastal
380 California grasslands: Effects of stand age and prescribed burning on control and
381 restoration. *Restoration Ecology*, *11*, 185–197.

382 Antonsen, H., & Olsson, P.A. (2005). Relative importance of burning, mowing and species
383 translocation in the restoration of a former boreal hayfield: responses of plant diversity and
384 the microbial community. *Journal of Applied Ecology*, *42*, 337–347.

385 Baeza, M.J., Luís, D., Raventós, J., & Escarre, A. (2002). Factors influencing fire behaviour
386 in shrublands of different stand ages and the implications for using prescribed burning to
387 reduce wildfire risk. *Journal of Environmental Management*, *65*, 199–208.

388 Bond, W.J., Woodward, F.I. & Midgley, G.F. (2005). The global distribution of ecosystems
389 in a world without fire. *New Phytologist* *165*, 525–538.

- 390 Castellnou, M., Kraus, D., & Miralles, M. (2010). Prescribed burning and suppression fire
391 techniques: from fuel to landscape management. In C. Montiel, & D. Kraus (Eds.), *Best*
392 *practices of fire use – prescribed burning and suppression fire programmes in selected*
393 *case-study regions in Europe* (pp. 3–16). European Forest Institute Research Report 24.
- 394 Cummings, D.C., Fuhlendorf, S.D., & Engle, D.M. (2007). Is altering grazing selectivity of
395 invasive forage species with patch-burning more effective than herbicide treatments?
396 *Rangeland Ecological Management*, 60, 253–260.
- 397 Davies, K.W., & Sheley, R.L. (2011). Promoting native vegetation and diversity in exotic
398 annual grass infestations. *Restoration Ecology*, 19, 159–165.
- 399 DiTomaso, J.M. (2000). Invasive weeds in rangelands: Species, impacts, and management.
400 *Weed Science*, 48, 255–265.
- 401 Fuhlendorf, S.D., & Engle, D.M. (2001). Restoring heterogeneity on rangelands: ecosystem
402 management based on evolutionary grazing patterns. *Bioscience*, 51, 625–632.
- 403 Fuhlendorf, S.D., Engle, D.M., Kerby, J., & Hamilton, R. (2009). Pyric herbivory: Rewilding
404 landscapes through the recoupling of fire and grazing. *Conservation Biology*, 23, 588–598.
- 405 Hansson, M., & Fogelfors, H. (2000). Management of a semi-natural grassland; results from
406 a 15-year-old experiment in southern Sweden. *Journal of Vegetation Science*, 11, 31–38.
- 407 Hobbs, N.T. (1996). Modification of ecosystems by ungulates. *Journal of Wildlife*
408 *Management*, 60, 695–713.
- 409 Howe, F.H. (1994). Response of early- and late-flowering plants to fire season in
410 experimental prairies. *Ecological Applications*, 4, 121–133.
- 411 Kahmen, S., Poschlod, P., & Schreiber, K.-F. (2002). Conservation management of
412 calcareous grasslands. Changes in plant species composition and response of functional
413 traits during 25 years. *Biological Conservation*, 104, 319–324.

- 414 Keeley, J.E. (2006). Fire management impacts on invasive plants in the western United
415 States. *Conservation Biology*, 20, 375–384.
- 416 Kiehl, K., Kirmer, A., Donath, T.W., Rasran, L., & Hölzel, N. (2010). Species introduction in
417 restoration projects – Evaluation of different techniques for the establishment of semi-
418 natural grasslands in Central and Northwestern Europe. *Basic and Applied Ecology*, 11,
419 285–299.
- 420 Köhler, B., Gigon, A., Edwards, P.J., Krüsi, B., Langenauer, R., Lüscher, A., & Ryser, P.
421 (2005). Changes in the species composition and conservation value of limestone
422 grasslands in Northern Switzerland after 22 years of contrasting managements.
423 *Perspectives in Plant Ecology, Evolution and Systematics*, 7, 51–67.
- 424 Liira, J., Issak, M., Jögar, Ü., Mändoja, M., & Zobel, M. (2009). Restoration management of
425 a floodplain meadow and its cost-effectiveness -- the results of a 6-year experiment.
426 *Annales Botanici Fennici*, 46, 397–408.
- 427 MacDougall, A.S., & Turkington, R. (2007). Does the type of disturbance matter when
428 restoring disturbance-dependent grasslands? *Restoration Ecology*, 15, 263–272.
- 429 Moog, D., Poschlod, P., Kahmen, S. & Schreiber, K.-F. (2002). Comparison of species
430 composition between different grassland management treatments after 25 years. *Applied*
431 *Vegetation Science*, 5, 99–106.
- 432 Page, H., & Goldammer, J.G. (2004). Prescribed burning in landscape management and
433 nature conservation: The first long-term pilot project in Germany in the Kaiserstuhl
434 viticulture area, Baden-Württemberg, Germany. *International Forest Fire News*, 30, 49–
435 58.
- 436 Parr, C.L., & Andersen, A.N. (2006). Patch mosaic burning for biodiversity conservation: a
437 critique of the pyrodiversity paradigm. *Conservation Biology*, 20, 1610–1619.

- 438 Pausas, J.G. (1999). Response of plant functional types to changes in the fire regime in
439 Mediterranean ecosystems: A simulation approach. *Journal of Vegetation Science*, 10,
440 717–722.
- 441 Pullin, A.S., & Stewart, G.B. (2006). Guidelines for systematic review in conservation and
442 environmental management. *Conservation Biology*, 20, 1647-1656.
- 443 Pyke, D.A., Brooks, M.L. & D'Antonio, C.M. (2010). Fire as a restoration tool: A decision
444 framework for predicting the control or enhancement of plants using fire. *Restoration*
445 *Ecology*, 18, 274-284.
- 446 Rietze, J. (2009). Ecological monitoring of the management of slope-vegetation by prescribed
447 burning in the Kaiserstuhl-Region, Germany. *International Forest Fire News*, 38, 63-67.
- 448 Rowe, H.I. (2010). Tricks of the trade: Techniques and opinions from 38 experts in tallgrass
449 prairie restoration. *Restoration Ecology*, 18, 253–262.
- 450 Ryser, P., Langenauer, R., & Gigon, A. (1995). Species richness and vegetation structure in a
451 limestone grassland after 15 years management with six biomass removal regimes. *Folia*
452 *Geobotanica & Phytotaxonomica*, 30, 157–167.
- 453 Twidwell, D., Fuhlendorf, S.D., Engle, D.M., & Taylor, C.A. (2012). Surface fuel sampling
454 strategies: linking fuel measurements and fire effects. *Rangeland Ecological Management*,
455 62, 223–229.
- 456 Vogels, J. (2009). Fire as a restoration tool in the Netherlands – First results from Dutch dune
457 areas indicate potential pitfalls and possibilities. *International Forest Fire News*, 38, 23–
458 35.
- 459 Wahlman, H., & Milberg, P. (2002). Management of semi-natural grassland vegetation:
460 evaluation of a long-term experiment in Southern Sweden. *Annales Botanici Fennici*, 39,
461 159–166.

462

462 Zusammenfassung

463 Graslandmanager und -wissenschaftler sind zunehmend an kostengünstigen Methoden des
464 Biodiversitätsschutzes auf Grasländern interessiert. Kontrolliertes Abbrennen ist eine
465 vielversprechende Methode, die bei der Planung von Managementmaßnahmen berücksichtigt
466 werden sollte. Darüber hinaus ist kleinräumiges kontrolliertes Abbrennen ein effektives Mittel gegen
467 die negativen Auswirkungen von Wildfeuern auf Ökosysteme und das menschliche Leben.
468 Kontrolliertes Abbrennen bildet einen integralen Bestandteil der praktischen
469 Graslandbewirtschaftung in Nordamerika, während es in Europa nur selten angewendet wird,
470 obwohl unkontrollierte Brände in manchen Regionen häufig auftreten. Unser Ziel war, den Nutzen
471 kontrollierten Abbrennens in Europa zu bewerten. Wir fanden, dass europäische Untersuchungen
472 zum kontrollierten Abbrennen auf Grasländern selten sind, und wir schließen, dass jährliches
473 Abbrennen gewöhnlich keine geeignete Option für den Schutz von artenreichen Grasländern
474 darstellt. Wir werteten auch Feuerstudien aus Nordamerika aus, um Befunde, die an die europäische
475 Strategie zum Graslandschutz angepasst werden könnten, zu identifizieren. In Nordamerika ist die
476 Anwendung von Feuer im Gegensatz zu Europa, was Häufigkeit und Zeitpunkt anlangt, fein
477 abgestimmt und normalerweise mit anderen Rekultivierungsmaßnahmen wie Beweidung oder
478 Aussaat kombiniert. Wir schließen somit, dass mit der Anwendung von sorgfältig geplantem
479 kontrolliertem Abbrennen zahlreiche Schutzziele (z.B. Kontrolle von invasiven Arten, Steigerung der
480 Landschaftsheterogenität) mit einer effektiven Feuerschutzstrategie verbunden werden können. Wir
481 betonen, dass für die Anwendung von kontrolliertem Abbrennen die allgemeinen Ergebnisse von
482 sorgfältig geplanten Fallstudien mit dem praktischen Wissen von Naturschutzmanagern über die
483 lokalen Anwendungsumstände kombiniert werden sollten, um spezifische Managementziele zu
484 erreichen.

485

485 Table 1. Summary of the results of prescribed burning studies in European grasslands.
 486

Country	Grassland and burning type	Positive effects of burning	Negative effects of burning	Recommendations	
Estonia	floodplain meadow, early spring burning, 4 times (within a 6 years' time)	reduction of litter, prevention of woody encroachment	species richness did not increase, contrary to mowing and mulching treatments	burning is not recommended	1 € 2
Germany	calcareous grassland, yearly winter burning (25 years)	reduction of litter	species composition was similar to that of fallow plots, cover of <i>Brachypodium pinnatum</i> increased	burning is not recommended	1 € 2
	calcareous grassland, yearly winter burning (25 years)	prevention of woody encroachment	species composition was similar to that of fallow plots, cover of <i>Brachypodium pinnatum</i> increased	burning is not recommended	1 € 2
	steppic grassland, late winter burning (1 year)	warmer and drier microclimate, delayed spread of existing trees	decline of snail individuals, the elimination of woody species was not complete	burning is recommended	1 € 2
	steppic grassland, late winter burning twice within a 4 years' time	most of the target species were not sensitive to burning	elimination of woody species and <i>Solidago gigantea</i> was not complete with solely burning	burning is feasible jointly with grazing or shrub clearance	1 2
Netherlands	dry dune grasslands, winter burning (1 year)	-	burning was not successful in nutrient removal	burning is feasible jointly with grazing	1 2
Sweden	commercial hayfield, early spring burning (1 year)	reduction of litter	species composition was similar to that of fallow plots	burning is not recommended	1 €
	semi-natural pasture, yearly early-spring burning (15 years)	reduction of litter	species richness declined, cover of tall herbs increased, untargeted species composition	burning is not recommended	1 €
	semi-natural grassland, yearly early spring burning (28 years)	-	species richness declined, untargeted species composition	burning is not recommended	1 €
Switzerland	limestone grassland, yearly winter burning (22 years)	increased cover of several rare plant species	species richness declined, cover of <i>Brachypodium pinnatum</i> increased	burning is not recommended	1 €
	limestone grassland, yearly winter burning (15 years)	reduction of litter	species richness declined, cover of <i>Brachypodium pinnatum</i> increased	burning is not recommended	1 €

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